




Article

Source Apportionment of Ground-Level Ozone and Assessment of Emission Reduction Strategies in Shenyang, China

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Abstract

This study presents an in-depth analysis of ground-level ozone (O₃) episodes in Shenyang from May to July 2019, utilizing advanced source apportionment modeling techniques: Ozone Source Apportionment Technology (OSAT), Geographic Ozone Assessment Technology (GOAT), and the High-order Decoupling Direct Method (HDDM). The research aimed to characterize the sources of O₃ and assess the potential impact of emission reduction strategies on O₃ concentrations. The results demonstrate that reducing emissions of both NO_x and VOCs can lower O₃ levels, with VOC controls proving to be more effective. During the ozone season, regional transport was identified as the dominant contributor to pollution, accounting for approximately 90% of the total, while local sources (sources within Shenyang's administrative boundary) contributed only about 10%. On days with severe pollution, the long-range transport of O₃ precursors was found to be the primary driver of Maximum Daily 8 h average O₃ (MDA8 O₃) exceedances in Shenyang. This study indicates that local measures alone are insufficient to eliminate O₃ exceedances; for instance, in some scenarios, even with a simultaneous 60% reduction in both local anthropogenic VOCs and NO_x emissions, MDA8 O₃ levels would not meet the national standard of 160 µg·m⁻³. Therefore, effective mitigation strategies must include regionally coordinated, time-dependent controls. This study also highlights that local industrial and mobile sources contribute to over 70% of ozone formation, suggesting that targeting these sectors could yield the most significant local benefits. This research underscores the need for a comprehensive approach to O₃ management, combining both local and regional efforts to address the complex issue of ground-level ozone pollution.

Keywords: ground-level ozone; source apportionment; VOC/NO_x control; regional transport; Shenyang



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1. Introduction

In recent years, China has made progress in controlling airborne particulate pollution [1–4], but surface O₃ pollution has become increasingly prominent as PM levels decline [5–8]. Shenyang, a core city of Northeast China, is vital to the regional revitalization plan, and its O₃ issue merits attention [9–11]. Data from Shenyang's national air-quality stations show that polluted days fell from 151 in 2013 to 63 in 2023, a drop of nearly 25%. While overall air quality has improved, O₃ has become the dominant pollutant on more of

these days. In 2013, O₃ was the primary pollutant on only 10 non-attainment days, but this peaked at 45 days in 2017. Although the figure fell to 12 days in 2021, it remained above 20 days in other recent years. Meanwhile, the 90th percentile of MDA8 O₃ stayed high, falling only slightly from 167 µg·m⁻³ in 2017 to 155 µg·m⁻³ in 2023, indicating that O₃ now limits further air-quality gains in Shenyang.

Recent field campaigns have significantly deepened our understanding of ground-level O₃ in Shenyang, with previous studies systematically examining its variability and meteorological drivers. During 2017–2022, the 90th percentile of MDA8 O₃ in urban Shenyang rose from 164 µg·m⁻³ to 192 µg·m⁻³ even though city-wide NO_x emissions decreased by 26% [12], a pattern attributable to enhanced ozone production efficiency under VOC-limited conditions. O₃ pollution in Shenyang exhibits pronounced seasonality, with summer peaks strongly driven by high temperatures, intense solar radiation, and the summer monsoon [13–15]. Liu et al. showed that under hot-dry-moderate-wind conditions, suburban O₃ production around Shenyang can outpace urban production by 20–30% [16]. Li et al. further found that a 10% rise in green-patch aggregation trims peak O₃ by 2–4 ppb under hot-stagnant conditions, revealing landscape-scale cooling and deposition feedbacks as key controls in Shenyang [17]. Collectively, the above studies establish that O₃ variability in Shenyang is governed not only by seasonal meteorology (temperature, humidity, monsoon circulation) but also by land-surface characteristics and precursor emissions [5,18–20]; disentangling and quantifying the relative contributions of meteorological drivers versus emission changes under mitigation scenarios therefore remains a critical, unresolved task.

Air-quality models can quantitatively assess how changes in emissions and meteorology affect regional O₃ concentrations and are therefore essential for identifying the sources of urban and regional O₃ [21–24]. The Comprehensive Air Quality Model with extensions (CAMx) is widely used for such simulations and for evaluating emission-control strategies [25–28]. In this study we applied the CAMx-OSAT module, the main tool for O₃ source apportionment, to quantify the contributions of regional and industrial sources to O₃ in Shenyang [29]. Previous CAMx-OSAT studies in Xining, Putian and Xinzhou have successfully identified the principal O₃ sources in those cities and provided scientific guidance for local controls [30–32]. To date, Shenyang-based O₃ research has relied mainly on observational analyses; integrated source apportionment and geographic O₃ assessment remain scarce. Yet northeast China borders the Beijing–Tianjin–Hebei megacity region, and pollutant transport between the two areas is substantial. Moreover, as a major gateway for China’s opening to the north and a hub for Northeast Asian cooperation, Shenyang’s air quality directly affects its international image. Therefore, we combine OSAT and GOAT to quantify the regional, industrial and geographic origins of O₃ in Shenyang, and use the high-order decoupled direct method (HDDM) to evaluate the O₃ response to precursor (NO_x and VOCs) emission reductions. The results will inform the design of effective O₃ control policies for the region.

2. Selection of Simulation Period

From 2013 to 2025, polluted days in Shenyang fell steadily. They peaked at 175 in 2014 (47.9% of the year) and dropped to 37 by 2025 (10.1%), underscoring Shenyang’s tangible progress in air-pollution prevention and control.

As total polluted days declined, O₃ became the dominant pollutant on an increasing share of non-attainment days (Figure 1). It should be noted that the primary pollutant was identified according to the calculation procedure prescribed by the National Ambient Air Quality Standards [33]. Only 10 such days were recorded in 2013, but this rose to 44 in 2017. Although the figure eased slightly in 2018–2019, it still accounted for ~40% of polluted

days, indicating that O₃ now limits further air-quality gains in Shenyang. In the years that followed, O₃ remained the dominant pollutant in Shenyang on roughly 40% of days, with only minor fluctuations tied to the exceptional circumstances of the COVID-19 period.

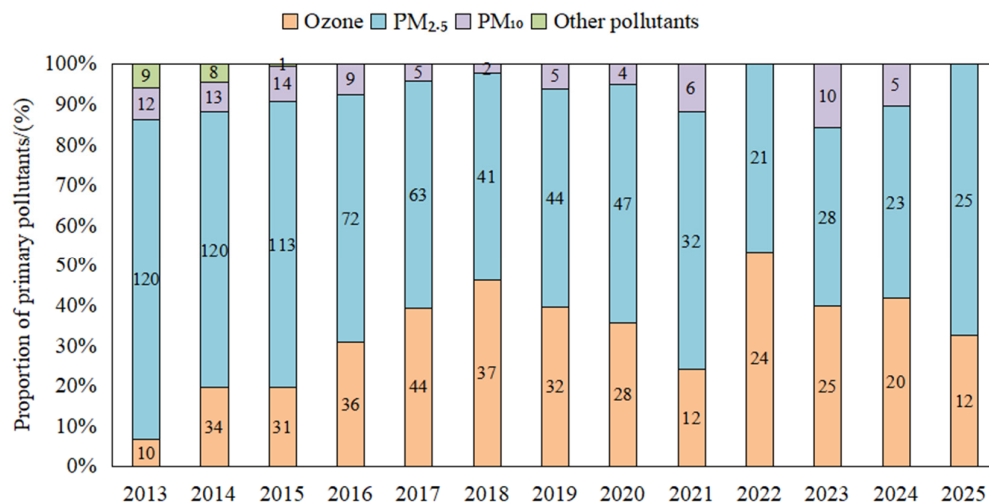


Figure 1. Trends in the number and proportion of days with different primary pollutants in Shenyang, 2013–2025.

Temperature and hours of sunshine are the main meteorological drivers of surface O₃ [13,14,34]. In Figure 2, Statistical analysis of exceedance days in Shenyang for 2013–2025 shows that high O₃ concentrations occur mainly in late spring and summer, especially from May to July. An extremely high MDA8 O₃ value, 262 µg·m⁻³, was recorded on 24 May 2019, and no such high MDA8 O₃ value had been observed in any of the subsequent years. Consequently, the WRF-CAMx model was applied to simulate O₃ concentrations in Shenyang for May–July 2019, defined here as the O₃ season. To quantify regional and sectoral contributions, we therefore focused on the heavy O₃ episode spanning 21–27 May 2019 that encompassed this peak (24 May 2019).

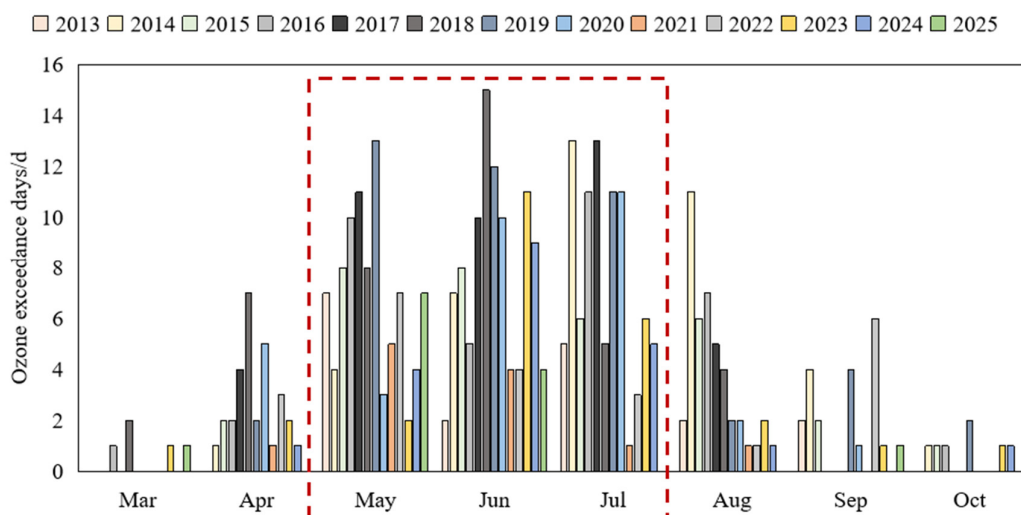


Figure 2. Monthly distribution of O₃ exceedance days in Shenyang, 2013–2025. Note: The months marked with red boxes are those with a higher number of ozone exceedance days.

3. Research Method

3.1. Model Setting

The simulation employed double-layer nested grids with the third-generation air-quality model CAMx v6.2 (Lambert conformal projection; central longitude 110° E, latitude 35° N). The outer domain (Domain 1) covered China (57–161° E, 1–59° N) at 36 km × 36 km resolution with 200 × 160 cells. The inner domain (Domain 2), nested at the 101st column and 87th row of Domain 1, encompassed Shenyang, Liaoning Province and the Beijing–Tianjin–Hebei region (109–127° E, 33–45° N) at 12 km × 12 km resolution with 119 × 101 cells. Twenty vertical layers extend to ~15 km. Meteorological fields were supplied by WRF v3.9 with a 1 h time step and grid spacing identical to CAMx. Emissions were derived from the MEIC 2019 inventory (0.25° × 0.25°, Tsinghua University), processed with the U.S. EPA’s Sparse Matrix Operator Kernel Emissions (SMOKE) model, and further refined with local Shenyang data for SO₂, NO_x, VOCs, PM_{2.5} and PM₁₀. Biogenic VOCs were computed with MEGAN. Gas-phase chemistry used SAPRC99 and aerosol chemistry the coarse-particle model; photolysis rates were calculated with the TUV model.

As shown in Figure 3, for the regional and geographic source analysis, the domain was divided into 20 source regions as listed below: Shenyang urban district (SYC), Kangping (KP), Faku (FK), Xinmin (XM), Liaozhong (LZ), Anshan (AS), Benxi (BX), Fushun (FS), Fuxin (FX), Jinzhou (JZ), Liaoyang (LY), Tieling (TL), Panjin (PJ), Huludao (HLD), Dandong (DD), Dalian (DL), Chaoyang (CY), Yingkou (YK), Beijing–Tianjin–Hebei (BTH), and other areas (OTH), that is, the remaining areas within the modeling domain. OSAT was then used to quantify the contributions of these regions to O₃ and its precursors (NO_x and VOCs), while the contributions of local industrial, residential, transport and power-plant emissions to NO_x, VOCs and O₃ in Shenyang were evaluated separately. Geographic O₃ formation was analyzed with GOAT. The receptor locations are the 10 national-grade air-quality monitoring stations operated within Shenyang’s urban districts, as listed below: Dongling Road (DL Rd), Xinxiu Street (XX St), Jingshen Street (JS St), Hunnan East Road (HN East Rd), Liaoshen West Road (LS West Rd), Yunong Road (YN Rd), Taiyuan Street (TY St), Xiaoheyuan (XHY), Wenhua Road (WH Rd), and Lingdong Street (LD St). The modeled concentrations were extracted from the corresponding grid cells, and their arithmetic mean was taken to represent the city-wide result for Shenyang.

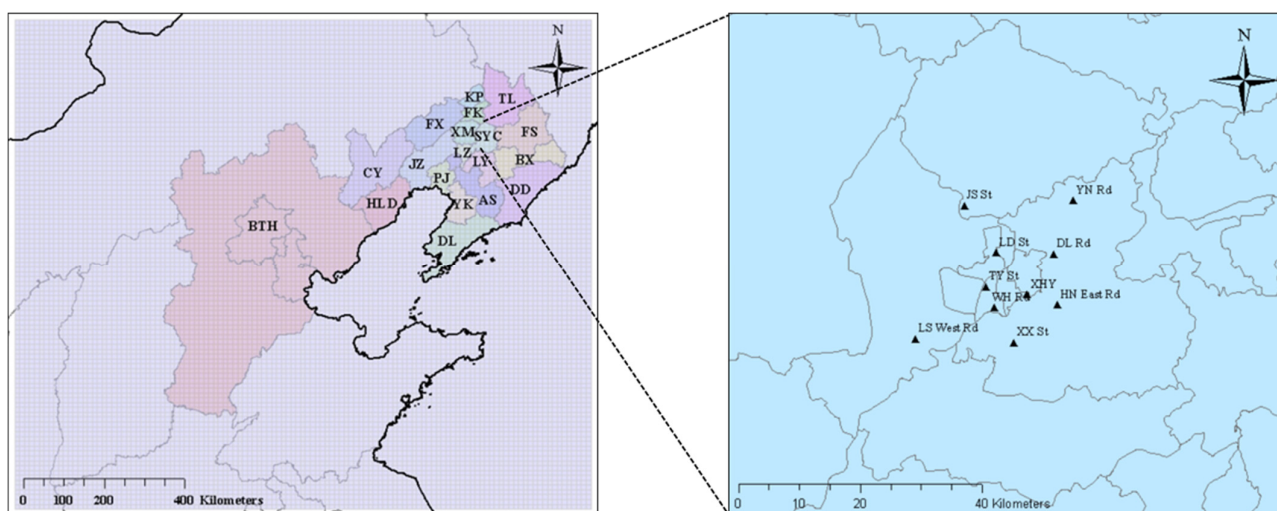


Figure 3. Schematic of the simulation domain, regional source regions and receptors.

3.2. Quantitative Analysis of Contributions

OSAT and GOAT embedded in the CAMx model were employed to quantify the contributions of different emission sources and geographical regions to ambient O₃ concentrations in the study domain.

OSAT employs a pollutant-tagging approach to quantify the contribution of O₃ precursors from different regions and source categories to O₃ formation. That is, OSAT tags the produced O₃, enabling the apportionment of contributions between locally generated and externally transported O₃. OSAT distinguishes between O₃ formed under NO_x-limited versus VOC-limited regimes, attributing production to the limiting reagent. The method provides source-specific O₃ contributions at each grid cell and timestep, enabling identification of dominant emission sectors and regional transport pathways.

GOAT extends OSAT by geographically tagging O₃ and its precursors according to their origin within user-defined receptor regions [35]. The technique employs a boundary-trajectory approach to track air mass transport, combining source region information with chemical apportionment. GOAT calculates the fractional contribution of each geographical source region to O₃ concentrations at designated receptor locations by integrating along backward trajectories. This method effectively separates local production from regional transport, quantifying trans-boundary O₃ contributions under varying meteorological conditions.

3.3. Assessment of Control Measures

Sensitivity analysis examines how the atmospheric system responds to changes in one or more variables and can quantify the relationship between ambient O₃ and emissions from different sources.

The decoupled direct method (DDM) is a widely used forward sensitivity analysis technique that computes semi-normalized sensitivity coefficients [36–38], as given by:

$$S_{ij}^* = \tilde{p}_j \frac{\partial C_i}{\partial p_j} = \tilde{p}_j \frac{\partial C_i}{\partial (\epsilon_j \tilde{p}_j)} = \frac{\partial C_i}{\partial \epsilon_j} \tag{1}$$

In the equation, \tilde{p}_j denotes the original source emissions, and ϵ_j denotes the fractional reduction (0–1).

The sensitivity parameters computed by DDM share the same units as the pollutant concentrations. DDM is more direct, efficient and stable than alternative sensitivity-calculation methods, and its sensitivity coefficients immediately reveal how emission reductions affect concentrations. For example, if $S_{ij} > 0$, cutting emissions from source j lowers the concentration of pollutant i ; if $S_{ij} < 0$, the reduction is ineffective or even raises concentrations.

Built on DDM, the high-order decoupled direct method (HDDM) computes first- and second-order sensitivity coefficients together with interaction terms [39]. Like DDM, its sensitivity parameters carry the same units as the pollutant concentrations. Once the sensitivity of concentration to emissions is known, the quantitative relationship between concentration and source control can be established via Taylor expansion; thus, the impacts of different NO_x and VOC control scenarios on ground-level O₃ in Shenyang are given by:

$$\Delta C_{O_3} \approx \Delta \epsilon_{NO_x} S_{NO_x}^{(1)} + \Delta \epsilon_{VOCs} S_{VOCs}^{(1)} + \frac{1}{2} \Delta \epsilon_{NO_x}^2 S_{NO_x}^{(2)} + \frac{1}{2} \Delta \epsilon_{VOCs}^2 S_{VOCs}^{(2)} + \Delta \epsilon_{NO_x} \Delta \epsilon_{VOCs} S_{NO_x VOCs}^{(2)} \tag{2}$$

4. Results and Discussion

4.1. Validation of Simulation Results

WRF was used to simulate and verify meteorological conditions (temperature, relative humidity and wind speed) in Shenyang from May to July 2019. Meteorological observations were obtained from Shenyang national meteorological station (WMO ID 54342), whose measurements are taken as representative of the city. WRF captures these variables well: temperature bias = -0.69 °C, RMSE = 2.55 °C, $R = 0.80$; relative-humidity bias = -1.71% , RMSE = 13.06% , $R = 0.68$. Near-surface wind speeds are overestimated, with bias = 2.04 m/s, RMSE = 2.40 m/s and $R = 0.56$. These findings agree with earlier studies reporting overestimated wind speeds in plains and valleys [40,41]. High wind speeds promote regional transport dominance by advecting upwind pollutants into the study area and diluting local emissions, potentially leading to an overestimation of exogenous O_3 contributions while masking weak local source signals in source apportionment results [42].

Based on CAMx, atmospheric pollutant concentrations in Shenyang from May to July 2019 were simulated and evaluated against national monitoring sites. As shown in Table 1 and Figure 4, CAMx reproduces ambient NO_2 and O_3 well, with correlation coefficients $R > 0.6$. Normalized mean bias (NMB) and normalized mean error (NME) for both species were within ± 0.5 and < 0.5 , respectively, indicating acceptable error [43]. Inventory uncertainties and the chemical mechanism are the main sources of NO_2 simulation error and also contribute to O_3 uncertainty [36]; additionally, the level of detail in the VOC emission inventory (spatial distribution, speciation, etc.) affects O_3 simulation accuracy.

Table 1. Model–observation comparison of NO_2 and O_3 concentrations in Shenyang, May–July 2019.

	O_3	NO_2
Monitored concentration ($\mu\text{g}/\text{m}^3$)	134.69	27.53
Simulated concentration ($\mu\text{g}/\text{m}^3$)	136.22	26.47
R	0.71	0.61
NMB	-0.003	-0.04
NME	0.23	0.22

Note: Observed NO_2 and O_3 data for Shenyang were provided by the Shenyang Academy of Environmental Sciences.

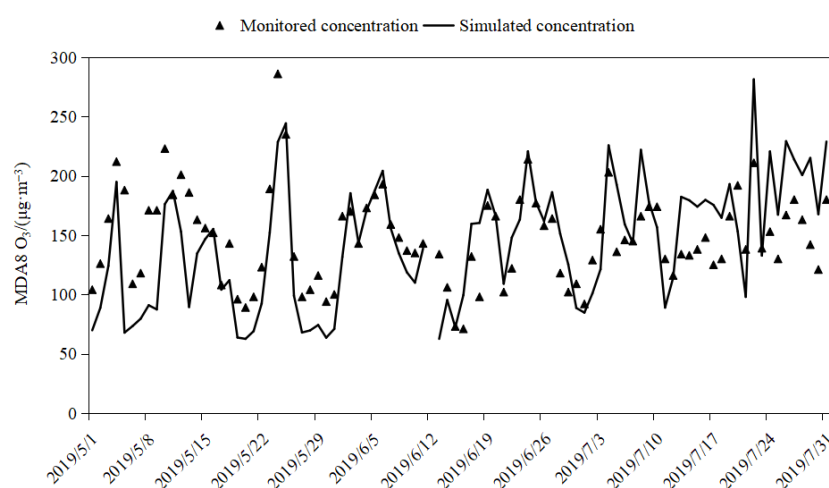


Figure 4. Cont.

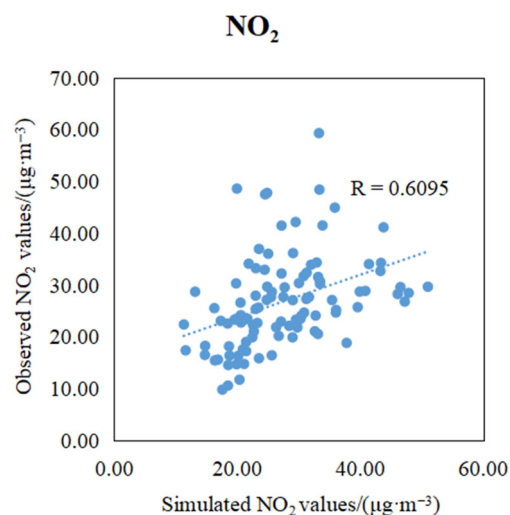


Figure 4. Comparisons of simulated MDA8 O₃ and NO₂ against observed data in Shenyang from May to July in 2019.

4.2. Identification of O₃ Control Zones in Shenyang

CAMx-DDM was used to quantify the sensitivity of ambient O₃ in Shenyang to NO_x and VOC emissions. Figure 5 shows that MDA8 O₃ in the north-west (KP, FK, XM) was positively sensitive to both precursors, but more strongly to NO_x, indicating a NO_x-limited regime; concurrent reductions are required, with NO_x control yielding the greater benefit. In contrast, south-east Shenyang (SYC, LZ) exhibited markedly higher sensitivity to VOCs than to NO_x, implying a VOC-limited regime where VOCs reductions are more effective. These patterns agree with Wang et al., Li et al. and Zhang et al. [44–46], who reported that Chinese urban and industrially developed areas are typically VOC-limited, whereas suburbs are NO_x-limited or transitional.

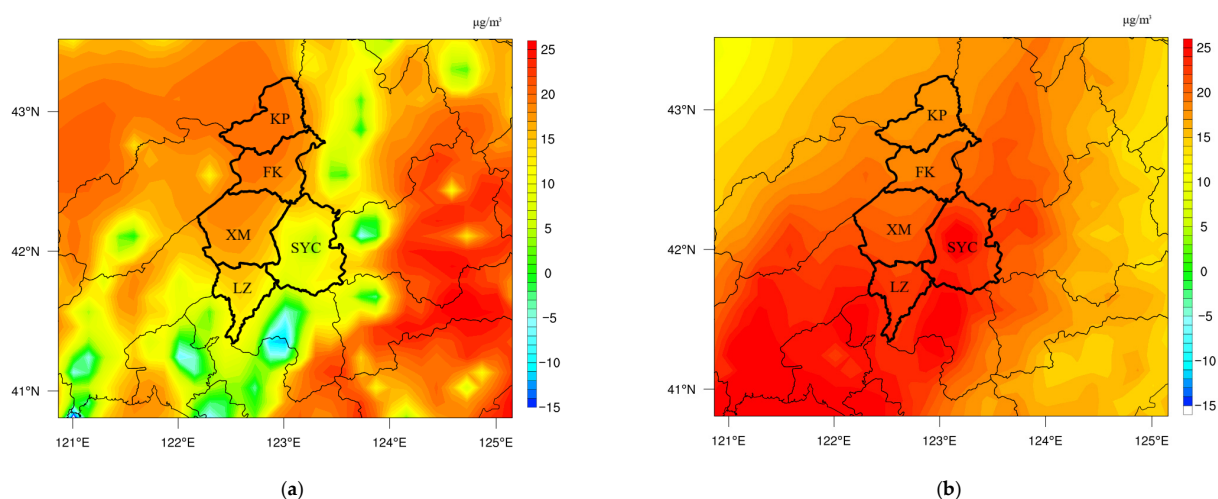


Figure 5. Sensitivity of O₃ to (a) NO_x and (b) VOC emissions in Shenyang and surrounding areas (120.8–125.2° E, 40.7–43.5° N).

Based on the above findings, we examined the hourly sensitivity of O₃ to VOCs and NO_x in Shenyang (averaged over the O₃ season). The reported values are the grid-cell averages extracted at the latitude–longitude positions of Shenyang’s national ambient air-quality monitoring stations. As shown in Figure 6, O₃ is more sensitive to VOCs; however, between 11:00 and 19:00 it is more sensitive to NO_x. Reducing NO_x emissions is therefore most effective during daytime, whereas VOC controls are more beneficial at other

times. Time-varying emission controls thus offer the greatest potential for O₃ mitigation. Consistent with previous CMAQ-HDDM-3D modeling studies over the Yangtze River Delta region that demonstrated a convex response of O₃ to NO_x emissions—with positive first-order and negative second-order sensitivities during afternoon peak hours, suggesting NO_x reduction is more effective than VOC-only control for mitigating peak O₃ [47]—this study further highlights the importance of time-specific and precursor-targeted emission control strategies for effective O₃ management.

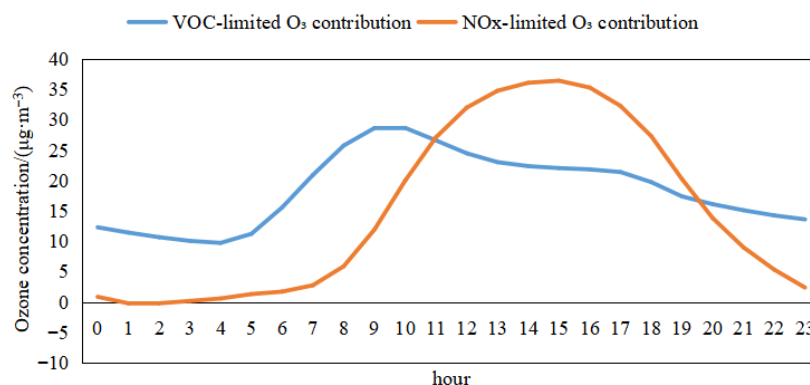


Figure 6. Time series of ozone contributions formed under NO_x-limited and VOC-limited conditions.

4.3. Regional and Local Industrial Contributions to O₃ Concentration

Figure 7 shows that more than half of the O₃ originated from other regions (OTH), according to the regional contribution simulation for O₃ in Shenyang (May–July 2019). Within Liaoning, Shenyang urban district (SYC) contributed 9.7%, and Anshan (AS) led at 8.0%, followed by Liaoyang (LY) and Dalian (DL); the remaining cities collectively added 12.1%. Beijing–Tianjin–Hebei (BTH) supplied 9.3%. It can be inferred from the above results that regional transport plays the predominant role. This can be attributed to the fact that regional transport not only elevates the local pollution baseline but also offsets local mitigation benefits through precursor advection, thereby fundamentally limiting the effectiveness of isolated urban control measures. Locally, industrial sources (IND) dominated at over 45%, natural emissions (NAT) contributed about 25%, power plants (POW) and transport (TRA) provided 16.0% and 13.3%, respectively, and residential (RES) sources had the smallest impact.

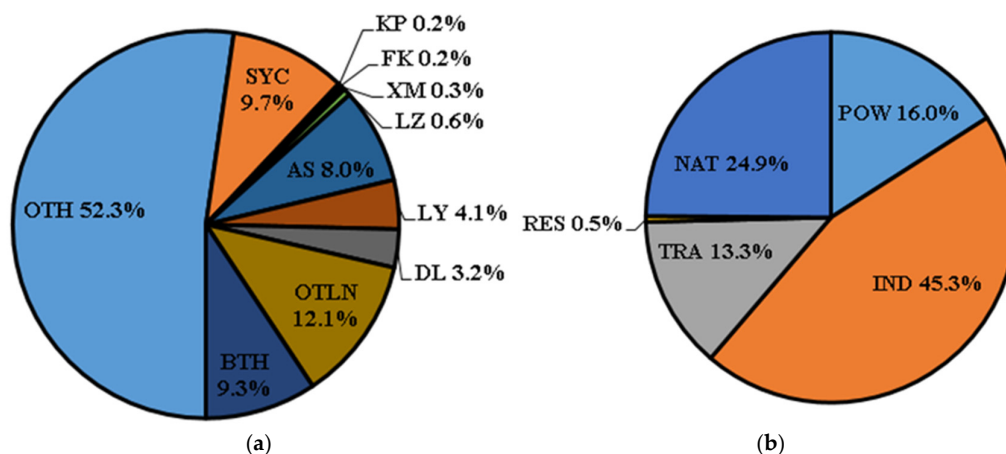


Figure 7. Regional (a) and local industrial (b) contributions to O₃ in Shenyang, May–July 2019.

4.4. Evaluation of Ozone-Precursor Emission Reductions

Based on the O₃ control zone results and the MDA8 O₃ data from Shenyang’s national air-quality monitoring sites, O₃ pollution days from May to July 2019 were classified into three levels:

Class III ($160 < \text{MDA8 O}_3 \leq 170 \mu\text{g}\cdot\text{m}^{-3}$, mean $166 \mu\text{g}\cdot\text{m}^{-3}$, slight), including 8 days;
 Class II ($170 < \text{MDA8 O}_3 \leq 200 \mu\text{g}\cdot\text{m}^{-3}$, mean $182 \mu\text{g}\cdot\text{m}^{-3}$, moderate), including 11 days;
 Class I ($\text{MDA8 O}_3 > 200 \mu\text{g}\cdot\text{m}^{-3}$, mean $218 \mu\text{g}\cdot\text{m}^{-3}$, severe), including 5 days.

The 90th percentile of the MDA8 O₃ for the same period was also analyzed. Emission reduction scenarios for NO_x and VOCs ranging from 10% to 60% (in 10% steps) were applied to evaluate the control effectiveness and the feasibility of attaining the O₃ standard in Shenyang.

Figure 8 and Table 2 show that reducing anthropogenic NO_x and VOC emissions in Shenyang markedly lowers ambient O₃, consistent with CMAQ-HDDM studies identifying northeast China’s spring–summer O₃ as a NO_x–VOC joint-control regime (Itahashi et al., 2013 [39]).

Table 2. O₃ response to local NO_x and VOCs emission reductions under varying pollution scenarios in Shenyang.

Emission Reduction Ratio		MDA8 O ₃ Concentration			
VOCs	NO _x	Class III	Class II	Class I	90th percentile
0	0	166	182	218	180
0	20	165	181	215	179
0	40	163	179	211	176
0	60	159	175	205	172
20	0	163	178	213	175
20	20	161	177	211	174
20	40	159	175	207	172
20	60	156	173	201	170
40	0	158	173	208	171
40	20	158	177	206	174
40	40	157	171	202	168
40	60	153	163	197	162
60	0	155	168	203	166
60	20	154	168	201	166
60	40	153	168	198	165
60	60	151	163	193	161

Note: Baseline values (0% reduction) are observation-adjusted measurements. Scenario values are calculated by multiplying baseline values with Relative Response Factors (RRFs) obtained from model simulations; The colors in the table are related to the magnitude of the values: larger values are displayed in red, while smaller values are shown in green.

Our analysis demonstrates that the efficacy of local emission controls is strongly dependent on O₃ pollution severity. Under moderate conditions (Class III), a 60% reduction in either VOCs or NO_x achieves the MDA8 O₃ standard ($160 \mu\text{g}/\text{m}^3$), with VOCs proving more effective; simultaneous 60% reductions yield an additional $\sim 11 \mu\text{g}/\text{m}^3$ decrease.

However, this approach fails under elevated pollution. For Class II and I scenarios, simultaneous 60% precursor reductions decrease MDA8 O₃ by $>20 \mu\text{g}/\text{m}^3$ and $\sim 25 \mu\text{g}/\text{m}^3$, respectively, indeed leading to measurable air-quality improvements, yet remain non-compliant. Similarly, at the 90th percentile, joint 60% reductions achieve only a $\sim 19 \mu\text{g}/\text{m}^3$ decrease, insufficient to meet standards. Thus, local controls alone are inadequate for severe O₃ events.

These limitations reflect the fundamentally regional nature of O₃ pollution, where transboundary precursor transport overwhelms local emission reductions during peak episodes. Consequently, regional joint prevention and control is essential for severe pollution management. Furthermore, the pronounced hourly variations in O₃ sensitivity to NO_x and VOCs necessitate time-resolved, precursor-specific mitigation strategies—such as staggered regional controls aligned with diurnal photochemical dynamics—to optimize efficacy and minimize socioeconomic costs. Future policies should integrate tiered responses: intensive local actions for routine conditions, escalating to coordinated regional interventions during severe episodes.

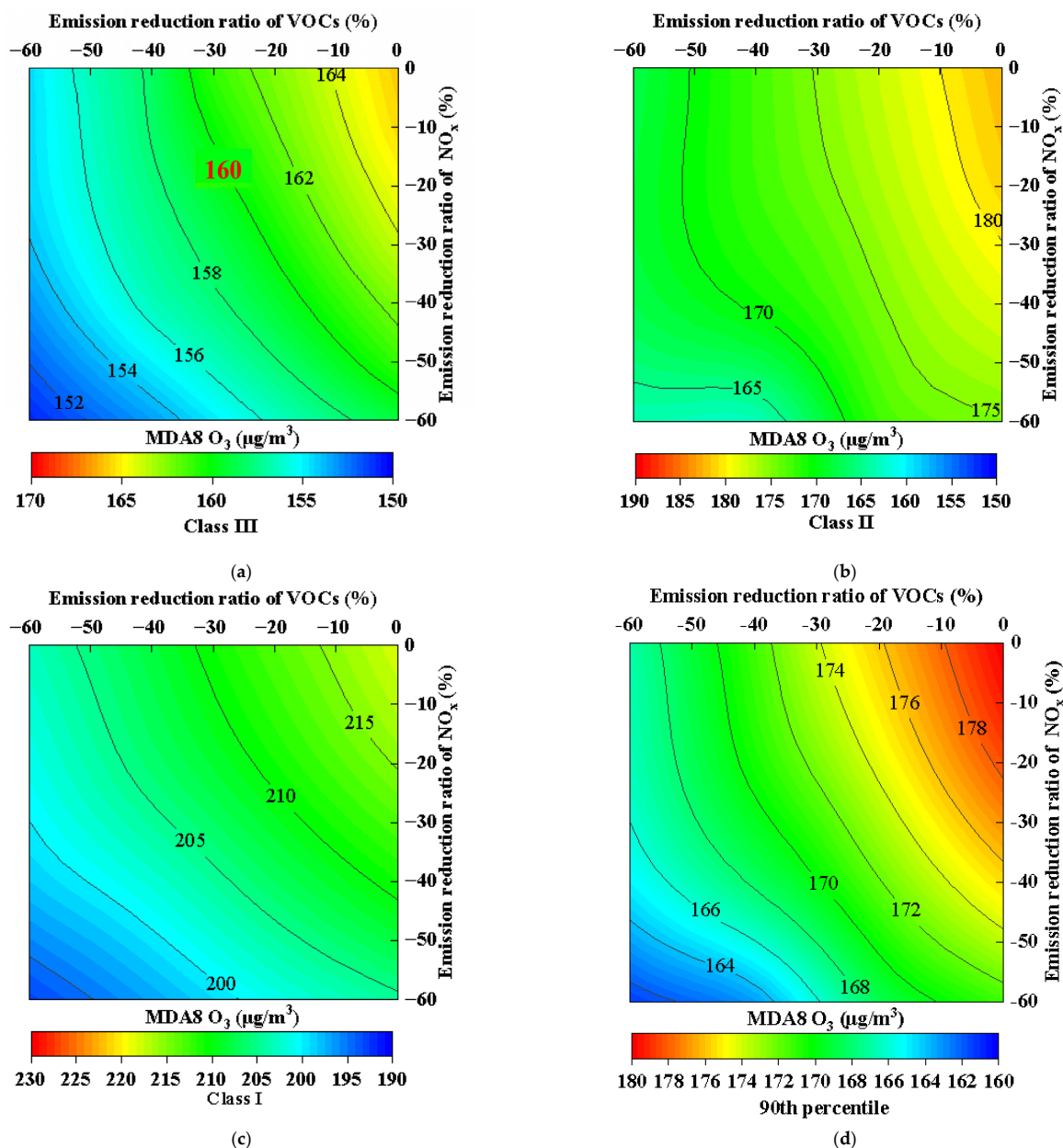


Figure 8. O₃ response to local NO_x and VOCs emission reductions under varying pollution scenarios in Shenyang. Note: The number 160 marked in red in the figure represents the secondary standard limit for O₃ concentration.

5. Analysis of Typical O₃ Pollution Process in Shenyang

5.1. Analysis of O₃ Pollution Attribution

As shown in Figure 9, no significant difference exists between OSAT and GOAT results for the mean O₃ pollution event in Shenyang during 21–27 May 2019. Both indicate that other regions (OTH) dominate, contributing ~60%. The key distinction is in source attribution: OSAT allocates more O₃ to long-range transport, whereas GOAT highlights local formation. Nevertheless, long-range transport remains a major driver during O₃ pollution episodes [35].

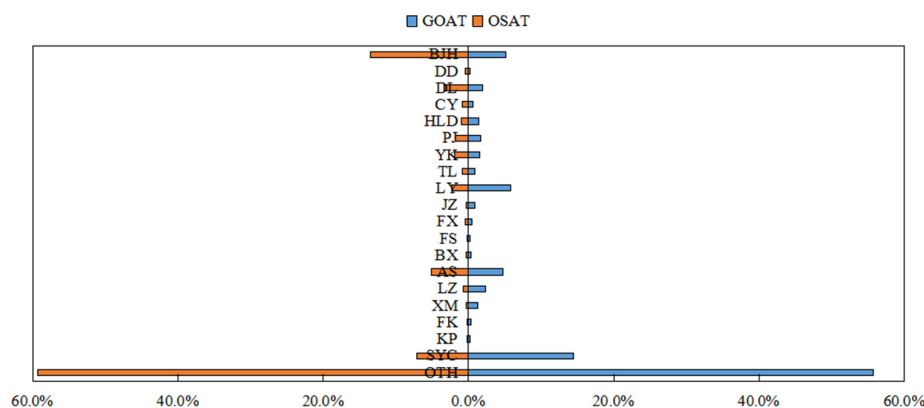


Figure 9. Comparison of OSAT and GOAT source apportionments during an O₃ pollution episode in Shenyang.

5.2. Industrial Contributions During the O₃ Pollution Episode

Figure 10 shows that during 21–27 May 2019, 78.8% of O₃ was formed under VOC-limited conditions, with industrial sources contributing the largest share—accounting for nearly 50% of total O₃ production, followed by natural sources. Meanwhile, 22.2% under NO_x-limited conditions, industry and power each contributed about 6% of total O₃ production. Therefore, during severe O₃ episodes, priority should be given to reducing large industrial VOC emissions (e.g., from coating, printing and dyeing operations).

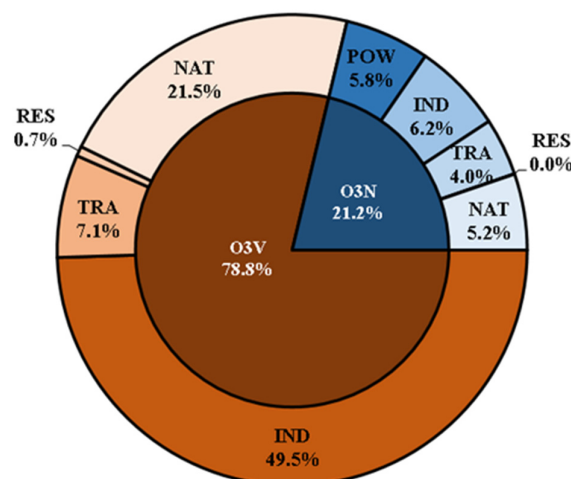


Figure 10. Sectoral contributions to MDA8 O₃ dominated by NO_x and VOCs in Shenyang, 21–27 May 2019.

6. Conclusions

This study conducts O₃ simulations for Shenyang during the O₃ season. O₃ source apportionment was used to quantify regional and sectoral contributions, and high-order

sensitivity analysis evaluated the effectiveness of various emission reduction scenarios. Shenyang lies in a NO_x–VOC collaborative control zone; simultaneous reductions in anthropogenic NO_x and VOCs alleviate O₃ pollution. Because O₃ sensitivity to these precursors exhibits pronounced hourly variations, time-varying controls are essential for effective mitigation.

Sectoral analysis shows that industrial VOC emissions—contributing nearly 50% of local O₃—should be prioritized for mitigation. However, O₃ in Shenyang is heavily influenced by regional transport of O₃ and its precursors, making it difficult to meet air-quality standards through local measures alone. While deep local emission reduction remains essential as a foundation for long-term air-quality improvement, regional joint prevention and control is the key to solving severe pollution episodes. Secondary pollutant management thus requires integrating local and regional strategies, particularly to address transboundary transport and chemical interactions.

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Informed Consent Statement: Not applicable.

Data Availability Statement: The original contributions to the study are included in the article; further inquiries can be directed to the corresponding author.

Conflicts of Interest: The authors declare no conflicts of interest.

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